Effect of temperature in removing of anions in solution on biochar using *Zea mays* stalks as a precursor

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Abstract: Biochar was prepared from corn (*Zea mays*) stalks and impregnated with sulfuric acid. The biomass was impregnated for 24 h with a 50% solution of H₂SO₄ with impregnation ratios 1:2 (B 1:2) and 1:3 p/v (B 1:3); then, it was carbonized in a muffle furnace at 520°C for 30 min with a 10°C per min ramp. The adsorption capacity to remove anions (nitrate, sulfate, and phosphate) in an aqueous solution was evaluated by varying the temperature. The adsorption mechanism was studied by determining the thermodynamic parameters: Gibbs free energy (∆G°), enthalpy (∆H°) and entropy (∆S°) standard. The biochars were characterized by Scanning Electron Microscopy-Energy Dispersive X-Ray Spectroscopy (SEM-EDS) analysis and were found to exhibit a heterogeneous surface and porous nature, with C, O, S, and Si. The experiments in the batch system showed the best performance of B 1: 2 in the removal of the three anions occurred at 303 K, while B 1: 3 had the best performance at 298 K. From the thermodynamic parameters, it was found that the removal processes are endothermic, their mechanism is by chemisorption. It is concluded that synthesized biochar is an excellent alternative to removing nutrient anions present in the solution.

Keywords: adsorption, biochar, cornstalks, nitrate, sulfate, phosphate

INTRODUCTION

The discharge of pollutants rich in nitrogen, sulfur, and phosphorus make higher contributions to the eutrophication of water [WURTSBAUGH et al. 2019]; with the entry of these substances into water bodies being the first step in triggering eutrophication in aquatic systems [STEBBINS et al. 2019]. Thus, eutrophication causes excessive enrichment of surface water with nutrients that promote autotrophic beings’ reproduction, mainly algae and cyanobacteria; this causes hypoxia or anoxia in poorly mixed waters [BEUSEKOM VAN 2017].

The entry of nitrates, sulfates, and phosphates into water bodies and their accumulation is attributed to anthropogenic sources: domestic and industrial wastewater, detergents, animal excrements, and fertilizers [BOEYKENS et al. 2017]. The intake of nitrate in high concentrations causes gastric diseases, congenital disabilities, cardiovascular diseases, thyroid damage, and vaccine cyanosis [WARD et al. 2018]. In addition, sulfate anions can lead to increased decomposition of organic matter, increased phosphate mobilization, and dissolved sulfur accumulation [FERNANDO et al. 2018].

Several conventional methods are used to remove water pollutants, such as membrane technologies, solar/UV degradation, filtration, electrochemical treatments and adsorption, among others [FAN, ZHANG 2018]. Bioadsorption is a frequently used technique because it allows the selective removal of specific solid components from a liquid phase through the use of biological material [FOMINA, GAID 2014], and its success due to the high availability of the precursors of the adsorbents, their reuse capacity, and the straightforward mechanism of the process [DAI et al. 2018].

Adsorbents for the removal of nitrate, phosphate, and sulfate have been developed using orange peel as precursor [KARTHIKEYAN, MEENAKSHI 2019], grapefruit [AKRAM et al. 2020],
peel and potato residues [El-Nahas et al. 2019], oil palm bagasse [Shang et al. 2018], corn residue [Wang et al. 2018], which reported good performance in anion removal after surface modification of the structure. Biochars are useful due to their porous structure, low cost, and environmental compatibility; however, it has shown a low adsorption capacity towards anionic species, making it necessary to modify the surface to improve its adsorption capacity. Thus, the objective of the present study was to evaluate the effect of temperature on the adsorption capacity of nitrate (N), sulfate (S), and phosphate (P), using a porous adsorbent prepared from functionalized corn stalks with different sulfuric acid impregnation ratios. The adsorption mechanism was studied by determining the thermodynamic parameters: Gibbs free energy ($\Delta G^o$), enthalpy change ($\Delta H^o$) and entropy change ($\Delta S^o$). The prepared biochar was characterized by SEM-EDS analysis to study the morphology and surface chemical composition. The pH zero charge point (pH$\text{PZC}$) was also determined to establish the pH value where the surface charge of the adsorbent is zero.

**STUDY METHODS**

In the experimental design, the adsorption capacity of phosphate, nitrate, and sulfate ions in mg·g$^{-1}$ was defined as a response variable. As the intervening variables, the pH, stirring rate (rpm), the concentration of the adsorbate (mg·dm$^{-3}$), and the dose of adsorbent (g). As an independent variable, the temperature in five levels: 298, 303, 308, 313, and 318 K. The synthetic solutions at 100 mg·dm$^{-3}$ were prepared using distilled water and mono-potassium phosphate (KH$_2$PO$_4$), sodium nitrate (NaNO$_3$), and potassium sulfate (K$_2$SO$_4$) of analytical grade, with 95% purity Merck Millipore brand. The biochar was activated with sulphuric ($\text{H}_2\text{SO}_4$) and hydrochloric acid (HCl) 1 M. The pH was adjusted with the H$^+$ and OH$^-$ ions [Bakatula et al. 2018].

**Preparation of biochar.** Corn stalks were collected as post-harvest waste in Maria la Baja (Bolivar, Colombia). The biomaterial was washed with deionized water, dried at 60°C for 12 h, and reduced in size in an electric mill. 50 g of pretreated corn stalks with a particle size of 1–2 mm were used. The biomass was impregnated for 24 h with a 50% solution of H$_2$SO$_4$ with impregnation ratios 1:2 (B 1:2) and 1:3 p/v (B 1:3); then, it was carbonized in a muffle furnace at 520°C for 30 min with a 10°C·min$^{-1}$ ramp. The biochar was washed with abundant distilled water until neutral. Finally, it was dried at 100°C for 8 h and stored in airtight containers until used in the adsorption tests [Mansanath, Kumar 2018]. The biofuels were characterized by analysis: Scanning Electron Microscope (SEM) in conjunction with an Energy Dispersion Spectrometer (EDS) to study the morphology and composition of the bio-adsorbent. The zero pH loading point (pH$\text{PZC}$) was determined to establish the pH value where the vast majority of the adsorbent sites are neutrally charged and therefore, the net charge (external and internal) of the particle is zero due to the reactivity of the substance’s surface with the H$^+$ and OH$^-$ ions [Barakula et al. 2018].

**Adsorption tests.** Adsorption tests were carried out in a batch system with phosphate, nitrate, and sulfate solutions prepared in synthetic solutions. The pH was adjusted with sodium hydroxide (NaOH) and hydrochloric acid (HCl) 1 M. The phosphate solution was prepared by dissolving 439 mg of dehydrated KH$_2$PO$_4$ in 1000 cm$^3$ of water [ASTM D 515-60]. To the nitrate one, KNO$_3$ was dried in an oven at 103–105°C for 24 h, it was dissolved 0.7218 g ±0.0005 g in deionized water and diluted up to 1000 cm$^3$. The sample was prepared with 2 cm$^3$ of chloroform per dm$^3$ [ASTM D7781-14]. The sulfate solution was prepared by dissolving 0.1479 g of dehydrated Na$_2$SO$_4$ in distilled water diluted to 1000 cm$^3$ [ASTM D 4130-15]. The tests were carried out in 15 cm$^3$ vials using an adsorbent dose of 2 g·dm$^{-3}$ at the optimum pH of each biochar, at 200 rpm on an orbital shaker at the set temperature conditions (298, 303, 308, 313 and 318 K).

Removal yield and adsorption capacity were determined according to Equations (1) and (2):

$$ RY = \frac{C_i - C_f}{C_i} \times 100 $$

$$ q_e = \frac{(C_i - C_f)V}{m} $$

where: $RY$ (%) is the removal efficiency, $C_i$ and $C_f$ are the initial and final concentration of the solution (mg·dm$^{-3}$), $q_e$ is the adsorption capacity of the biochar (mg·g$^{-1}$), $V$ is the volume of solution and (dm$^3$), $m$ is the mass of adsorbent used (g).

**Thermodynamic parameters.** Thermodynamic parameters such as Gibbs' free energy ($\Delta G^o$), enthalpy change ($\Delta H^o$), and entropy change ($\Delta S^o$) were determined to learn about the nature of adsorption and its thermodynamic behaviour. This fact was done by the Van't Hoff graphical method, summarized in the following equations [Dobrosz-Gómez et al. 2018]:

$$ \Delta G^o = -RT\ln(K_c) $$

$$ \ln(K_c) = -\frac{\Delta H^o}{RT} + \frac{\Delta S^o}{R} $$

where: $R$ is the ideal gas constant and has a value of 8.314 J·mol$^{-1}$·K$^{-1}$, $T$ is the temperature expressed in Kelvin, $K_c$ is the adsorption equilibrium constant.

**RESULTS AND DISCUSSION**

**CHARACTERISATION OF BIOCHAR**

As shown in Figure 1, the bio-adsorbents exhibited a heterogeneous surface and porous nature. The B 1:3 ratio was that of the most porous exposed area, attributed to the higher proportion of H$_2$SO$_4$ used during the material’s impregnation. The pores are in the macro and mesopore range and allow access to the microporous structure, similar to that reported by [Par et al. 2018]. B 1:2 shows the presence of agglutinated networks and a lower presence of pores. The graphitized structure and the degree of aromaticity are affected by the volatilization of oxygen and hydrogen from the lignocellulosic compounds of low aromaticity and lignin of the corn stalks. This fact caused the remaining carbon to form new aromatic bonds [Peiris et al. 2019].
From the EDS images, it was observed the presence of C, O, Si, and S atoms, related to the bio-adsorbents; the sulphur related to the impregnation of the biochar with H$_2$SO$_4$ evidence the success of the process, considering that the S content of the biochar without impregnation was 2.35% w/w and in the corn stalks 4.3% w/w. A summary of the elemental composition determined by EDS analysis is shown in Table 1.

Figure 2 shows a graph of the final pH versus the initial pH for the differently prepared biochar and calculates the pH$_{pzc}$. Biochar behaves like a positively charged structure with pH values below pH$_{pzc}$. After increasing the solution’s pH, the bio-adsorbents behave similarly because of the deprotonation of their surface [HASSAN et al. 2017]. The pH$_{pzc}$ of B 1:2 was 2.95, and B 1:3 was 3.65, indicating that pH values above this value should have a negatively charged surface [IZQUIERDO 2010]. Different authors have reported similar values for the pH$_{pzc}$ in carbons activated with sulfuric acid using precursors of different biomaterials of lignocellulosic origin: PEIRIS et al. [2019] found a pH$_{pzc}$ of 3.5, while 4.56 was reported by PAN et al. [2018] found this property at 4.5. It is observed that the range of the reported values is between 2 and 5. The synthesized biochar has acid groups on its surface due to the impregnation with H$_2$SO$_4$ at 50% v/v. The low pH of B 1:2 and B 1:3 can be attributed to increased hydron cations or the decrease of hydroxyl anions [FAN, ZHANG 2018]. Considering that the synthesized biofuels will be used to remove anions, it is considered that by using the pHs of the solution below the pH$_{pzc}$, they would positively charge the material’s surface, promoting the adsorption of the solution anionic contaminant. Thus, it was established that the optimum operating pH for B 1:2 was 2.5, and for B 1:3, it was 3.5.

**EFFECT OF TEMPERATURE**

The operational pH for the bio-adsorption process of each anion with the biochar has been defined (2.5 and 3.5 for B 1:2 and B 1:3, respectively). The experiments were run at varying temperatures. Figure 3 summarises the results of the effect of temperature on the removal capacity of nitrate (N), sulfate (N), and phosphate (P). The removal best performance of B 1:2 was at 303 K, while B 1:3 had the best performance at 298 K. Therefore, it is assumed that the process does not need to be supplied with energy in order for this to happen due to the possible exothermic nature of the reactions. The bio-adsorbents showed selectivity by the phosphate, obtaining the highest adsorption with B 1:2, due to the additional H$^+$ adsorption of the solution on the biochar surface [ZARE, GHASEMI-FASAEI 2018].

Table 2 summarises the thermodynamic parameters of nitrate, sulfate, and phosphate adsorption; it is determined that all removal processes present endothermic adsorption due to the negative sign of ΔH°. Thus there is no necessity to supply energy to the system during the process [BASHIR et al. 2017]. According to its value, chemisorption is suggested for the

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**Table 1. Composition of biochar (B)**

<table>
<thead>
<tr>
<th>Element</th>
<th>Percentage by weight (% w/w)</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>B 1:2</td>
</tr>
<tr>
<td>C</td>
<td>75.86</td>
</tr>
<tr>
<td>O</td>
<td>11.75</td>
</tr>
<tr>
<td>Mg</td>
<td>–</td>
</tr>
<tr>
<td>Si</td>
<td>1.67</td>
</tr>
<tr>
<td>P</td>
<td>–</td>
</tr>
<tr>
<td>S</td>
<td>10.72</td>
</tr>
<tr>
<td>K</td>
<td>–</td>
</tr>
<tr>
<td>Ca</td>
<td>–</td>
</tr>
</tbody>
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Source: own study.
processes [SERESHTI et al. 2020]. The positive values of $\Delta G^\circ$ indicate that the system is not spontaneous for removing anions using the synthesized biochars. Furthermore, the $\Delta G^\circ$ increases as the system's temperature increases, the process energetically becomes more favourable [MEHDINEJADIANI et al. 2019]. The negative value of $\Delta S^\circ$ indicates the removal irreversibility and a low probability of structural changes due to the formation of the bonds between the active centres of the bio-adsorbent and the pollutants [YIN et al. 2019].

**CONCLUSIONS**

1. B 1:3 has a porous structure with macro and mesopores, and B 1:2 show the presence of agglutinated with the presence of pores.

2. The best performance of B 1:2 in the removal of the three anions was at 303, while B 1: 3 had the best performance at 298 K.

3. Thermodynamic parameters established that the removal processes of the three anions are endothermic. Their mechanism is by chemisorption, the adsorption is favourable with a probability of structural changes by the formation of the bonds between the active centres of the bio-adsorbent and the pollutants. From the results, it is determined that biochar is a good alternative for removing nutrient anions present in the solution.

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**REFERENCES**

